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Plant-soil feedbacks of exotic plant species across life forms: a meta-analysis

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Meisner et al.

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Meisner et al.

35 Abstract (150-250 words)

Invasive exotic plant species effects on soil biota and processes in their new range can promote or counteract invasions via changed plant-soil feedback interactions to themselves or to native plant species. Recent meta-analyses revealed that soil influenced by native and exotic plant species is affecting growth and performance of natives more strongly than exotics. However, the question is how uniform these responses are across contrasting life forms. Here, we test the hypothesis that life form matters for effects on soil and plant-soil feedback.

In a meta-analysis we show that exotics enhanced C cycling, numbers of meso-42 invertebrates and nematodes, while having variable effects on other soil biota and processes. 43 Plant effects on soil biota and processes were not dependent on life form, but patterns in 44 feedback effects of natives and exotics were dependent on life form. Native grasses and forbs 45 caused changes in soil that subsequently negatively affected their biomass, whereas native trees 46 caused changes in soil that subsequently positively affected their biomass. Most exotics had 47 neutral feedback effects, although exotic forbs had positive feedback effects. Effects of exotics 48 on natives differed among plant life forms. Native trees were inhibited in soils conditioned by 49 exotics, whereas native grasses were positively influenced in soil conditioned by exotics. We 50 conclude that plant life form matters when comparing plant-soil feedback effects both within 51 52 and between natives and exotics. We propose that impact analyses of exotic plant species on the performance of native plant species can be improved by comparing responses within plant 53 life form. 54

55

Keywords: alien plant species, exotic plant species, life form, meta-analysis, plant invasions,
plant-soil feedback, plant-soil interactions, soil legacies

Meisner et al.

58 Introduction

Plants can affect abiotic and biotic soil properties causing feedback interactions to themselves, 59 their offspring, or to (the offspring of) other plant species (Wardle et al. 2004; Ehrenfeld et al. 60 61 2005; Bever et al. 2010) (Fig. 1). An increasing number of studies suggest that the abundance of exotic plant species may be influenced by them altering soil conditions in a manner that 62 benefits their own performance through positive feedbacks (Callaway et al. 2004; Reinhart and 63 Callaway 2004; Agrawal et al. 2005; Engelkes et al. 2008; Maron et al. 2014), which may 64 provide them with a competitive advantage in their new range. These suggestions are generally 65 confirmed by recent meta-analyses (Kulmatiski et al. 2008; Suding et al. 2013). However, little 66 is known about how uniform these plant-soil feedback interactions are across life forms both 67 within and between native and exotic plant species. 68

Overall, native plant species experience variable, but predominantly negative plant-soil 69 feedbacks (Reinhart 2012; Mangan et al. 2010; Fitzsimons and Miller 2010; McCarthy-70 Neumann and Kobe 2010; Kulmatiski et al. 2008), whereas introduced exotics generally 71 experience neutral or even positive plant-soil feedbacks (Callaway et al. 2004; Reinhart and 72 Callaway 2006; Suding et al. 2013; Engelkes et al. 2008). The magnitude of plant-soil 73 feedback effects for plant species in greenhouse studies has been observed to correlate with the 74 abundance of plant species in the field (Klironomos 2002; Mangan et al. 2010; McCarthy-75 Neumann and Ibáñez 2013). These findings lead to the conclusion that invasiveness of 76 introduced exotic plant species is because they are subjected to less negative feedback with soil 77 than are native plant species. However, the correlation between the magnitude of plant-soil 78 feedback and plant species abundance in the field is not observed in all studies (Reinhart 2012) 79 and there have been few empirical tests under field conditions (Casper and Castelli 2007). In 80 addition, only a small portion of introduced exotic plant species become invasive (Williamson 81 and Fitter 1996). Moreover, exotics may also indirectly benefit from altered plant-soil feedback 82 when disturbing the positive feedback effect of some native plant species (Suding et al. 2013). 83

101

Meisner et al.

For example, because some exotics reduce symbiotic mutualists (Stinson et al. 2006), increase
local pathogens (Eppinga et al. 2006) or possibly accumulate allelochemicals (Callaway and
Ridenour 2004).

87 In many studies effects of introduced exotic plant species on soil have been measured in relation to changes in specific soil biota (Stinson et al. 2006; Vogelsang and Bever 2009) and 88 soil processes (Vilà et al. 2011; Kourtev et al. 2003; Meisner et al. 2012). Some of these 89 studies use experimental data, whereas other studies are based on observational differences 90 between uninvaded versus invaded areas. The use of experimental data has an advantage as it 91 enables separation of causes and consequences, but the short duration of most experiments has 92 93 a disadvantage in that not all soil factors may have had sufficient time to respond to the presence of the exotics. The main disadvantage of observational data is that the observed 94 effects may have been the cause of invasiveness, rather than the consequence. Ideally both 95 observational and empirical studies should run in parallel or need to complement each other. 96 Feedback effects may be dependent on plant species, taxonomic group or life form. For 97 example, grasses and forbs have in general a more negative feedback than trees (Kulmatiski et 98 al. 2008). Thus far, it is unknown if exotic and native species differ in plant-soil feedbacks 99 across plant life form (Liao et al. 2008; Suding et al. 2013). Therefore, in our meta-analysis, we 100

102 effects within plant life form: trees, forbs, grasses, and nitrogen (N)-fixing plant species.

studied effects of both exotic and native species on soil properties and plant-soil feedback

We first explored the effects of exotics on specific groups of soil biota and soil processes, in order to advance beyond the black-box approach of plant-soil feedback (Cortois and De Deyn 2012; Van der Heijden et al. 2008; Bever et al. 2010). Then, we explored plantsoil feedback differences in the response of exotic and native plant species to soil conditioned by either the exotic or native species. In this way our meta-analysis, complements the analysis of Vilà et al. (2011) and addresses different questions than the analyses performed by Kulmatiski et al. (2008) and Suding et al. (2013). We determined if the soil-mediated

Meisner et al.

feedbacks of exotics and natives to themselves and other groups of plants (exotics to natives and natives to exotics) differed across plant life forms. We tested the hypotheses that: (1) exotic plant species will enhance process rates and promote soil biota; (2) exotics experience less negative plant-soil feedback from their own soil than do co-occurring native species; (3) native species experience a more negative feedback from soil influenced by exotic species than vice versa. For each hypothesis, we tested to what extent the outcome depended on plant life form.

117

118 Material and methods

119 *Literature search*

Literature was searched using Web of Science and Scopus with combinations of the following
keywords: exotic plant, introduced plant, rhizosphere, invasi* plant, biota, soil, litter, feedback,
priority effect, soil legacies. Papers were also selected based on references in other papers and
cited papers. A total of 203 papers were selected to screen if the data fitted the inclusion
criteria.

125

126 Inclusion criteria for effects of exotics on soil biota and soil processes

We studied effects of exotics on soil biota and soil processes using soils conditioned by exotic 127 species as treatments and soil conditioned by native species as controls. We used only studies 128 in which the compared exotic and native species were co-occurring in the new range of the 129 exotic plant species. We evaluated effects of both rhizosphere and litter inputs on soil (Fig. 1). 130 If the experiment was performed in both invaded and non-invaded soils, we only used data 131 from the non-invaded soils to determine the effect size of exotic species before entering the 132 new range. We included studies that collected rhizosphere or litter from the field. The type of 133 comparison was noted: with native species (same life form, other life form or congener), plant 134 input (rhizosphere or litter), and type of study (field or greenhouse). Effects of exotics were 135

Meisner et al.

specified to: AMF, fungal biomass, bacterial biomass, microbial biomass, invertebrate count,
nematode count, C cycle, N cycle or P cycle. Supplemental Table S1 presents measurements
included within the different categories.

139

140 Inclusion criteria for feedbacks from their own soil

Here we compared feedbacks of native and exotic species in soil conditioned by conspecifics 141 (own soil). We calculated effect sizes by considering own soil as the treatment and away soil 142 (unconditioned soil, soil conditioned by congeners, soil conditioned by other species or 143 sterilized soil) as the control. Only those studies were included where exotic and native species 144 in the experiment co-occurred in the new range of the exotic plant species. The method used to 145 determine plant biomass was recorded: aboveground biomass, total biomass, or other biomass 146 measure. Native and exotic species were classified according to life forms (grass, forb, tree, N-147 fixing). One specific nutrient acquisition trait (N-fixing) was added, because this trait may 148 relate to invasiveness (Liao et al. 2008). The studies that met the inclusion criteria are 149 150 presented in Table S2.

151

152 Inclusion criteria for feedbacks of exotics to natives and natives to exotics

We compared feedback of exotics to natives and of natives to exotics by considering performance in away soil as treatment and in own soil as control. Away soil of natives was conditioned by the exotics and away soil of exotics was conditioned by natives. As above we noted the method to determine biomass in each study, and the life form of each native and exotic species. The studies that met the inclusion criteria are presented in Table S3.

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159 Data extraction and calculating effect sizes

When data met the inclusion criteria, means, variance estimates (SE or SD) and number of
replicates (n) were extracted. Out of the 203 papers, we selected 30 papers on feedback effects

Meisner et al.

of home soil, 32 papers on feedback effect of native to exotics and exotic to natives, and 39 162 papers on effects of exotics on soil biota and soil processes (see supplemental information). For 163 papers with multiple plant pairs, we considered plant species as unit of replication (Gurevitch 164 165 et al. 2001). We extracted data (means and variance estimates) from graphs with DataThief (B. Tummers, DataThief III. 2006 http://datathief.org/). When data (mean, variance and/or n) was 166 missing from the study, data were obtained via contacting the corresponding authors of papers. 167 Studies with authors that could not be traced were omitted. To avoid non-independence, we 168 calculated a pooled mean and a pooled standard deviation for the treatment or control when 169 there was more than one treatment and only one control. We also did this for the treatments 170 171 when there was more than one control and only one treatment (Borenstein et al. 2009; Van Kleunen et al. 2010). 172

173

174 *Calculating effect sizes*

For each parameter of interest, a standardized mean effect size per species was determined by 175 calculating Hedges'd using Metawin 2.0 (Rosenberg et al. 1999). This is the standardized mean 176 difference between the treatment and the control that is weighted by the pooled variance 177 (Borenstein et al. 2009; Gurevitch and Hedges 2001) and multiplied by factor J to correct for 178 bias of small sample size (Gurevitch and Hedges 2001; Rosenberg et al. 1999). These 179 individual effect sizes were combined by calculating a pooled summary effect size over all 180 species for each of the parameters of interest using a random model. A random model is 181 appropriate for ecological data as this takes heterogeneity between species into account 182 (Borenstein et al. 2009; Gurevitch et al. 2001). We calculated bias-corrected 95 % bootstrap 183 confidence intervals using 4999 iterations (Adams et al. 1997). Effect sizes were significantly 184 positive or negative when these confidence intervals did not overlap with 0 at P < 0.05 and the 185 sign of the effect size relates to positive and negative feedback, respectively. For effects of 186

Meisner et al.

exotics on soil biota and processes, a positive effect size indicated that exotics increased thesoil parameter of interest, while a decrease was indicated by a negative effect size.

We tested the variation between the effect sizes using a homogeneity test (O), which 189 190 was evaluated using a chi-square test of significance. This test evaluates the null hypothesis that all studies share the summary effect size (Borenstein et al. 2009). When Q_{total} is 191 significant, it indicates that effect sizes are not equally distributed across the studies in the 192 meta-analysis, or that the direction of effect sizes varies between studies. Provided that 193 sufficient data were available, we calculated the effect sizes per category of origin (native 194 versus exotics), biomass measurement type or life form. We tested if the direction of effect 195 sizes differed between categories (Q_{between}) and the extent to which effect sizes contained 196 variation that was unexplained by categories (Q_{within}). 197

198

199 *Corrections for non-independence of effect sizes*

Some plant species within studies contained more than one effect size, such as when pots were 200 sampled at multiple time points or when studies were performed in soil from multiple 201 locations. Effect sizes within a study were combined by calculating the fixed summary effect 202 size and variation for each plant species to avoid non-independence of the effect sizes, when 203 204 measurements were from multiple experiments within a study, such as multiple environments or multiple independent time points, (Borenstein et al. 2009; Van Kleunen et al. 2010). When 205 measurements were performed over a time course, we used effect sizes from the final sampling 206 date. When there were more than one measurement on one individual plant species (e.g. two 207 kinds of C cycling measurements), we combined the data as described in chapter 24 of 208 Borenstein et al. (2009) and calculated a pooled mean effect size for the effect sizes from the 209 different measurements. The pooled variation of the mean effect size was calculated with the 210 211 following formula:

$$var\left(\overline{Y}_{i}\right) = \left(\frac{1}{m}\right)^{2} var\left(\sum_{i=1}^{m} Y_{i}\right) = \left(\frac{1}{m}\right)^{2} var\left(\sum_{i=1}^{m} V_{i} + \sum_{i \neq j} \left(r_{ij} \sqrt{V_{i}} \sqrt{V_{j}}\right)\right)$$

where *var* (\bar{Y}_i) is the pooled mean variance of effect size \bar{Y}_i for *m* variables. The correlation coefficient r_{ij} describes to which extent \bar{Y}_i and \bar{Y}_j co-vary, but *r* is often unknown. When *r* = 1, the variances are completely dependent on the different measurements and when *r* = 0, the variances are completely independent. The variances will affect the relative weight of the effect size when calculating the summary effect size with more weight going to the study with lower variance. We used *r* = 1 as this is the most conservative approach (Davidson et al. 2011) and we obtained similar results when *r* = 0.

219

220 *Checking for bias in data*

We calculated Rosenthal's fail safe numbers to address the "file drawer problem", which is the 221 problem that studies with strong treatment effects are more likely to be published than studies 222 223 with no or weaker treatment effects (Borenstein et al. 2009). Thereto, we calculated the number of studies needed to change the outcome of a significant summary effect size to non-224 significant. Fail safe numbers should be approximately larger than 5n + 10 where n = number 225 of studies. We also performed a rank correlation test, Spearman Rho, between effect size and 226 variance. A significant correlation indicates that larger effect sizes in one direction are more 227 likely published than smaller effect sizes (Rosenberg et al. 1999). We inspected data visually 228 for abnormalities in data structure that would indicate publication bias by drawing a funnel plot 229 and a Normal Quantile Plot. 230

231

- 232 **Results**
- 233 *Effects of exotics on soil biota and processes*

Exotics had positive effects on invertebrate abundance, nematode abundance and the processes

involved in the carbon cycling (Fig. 2, see Table S1 for processes measured in experiments).

Meisner et al.

Effect sizes for effects of exotics on AMF, P cycling and N cycling differed between studies 236 (Q_{total} in Table S4), meaning that effect sizes were positive, negative and neutral depending on 237 the study. The comparison with natives (same life form, other life form, or congener), plant 238 239 input (rhizosphere or litter inputs) or type of study (field, greenhouse) could not explain the differences in effect sizes, as indicated by non-significant values of Q_{between} (P > 0.1). There 240 may be a bias in the effects of exotics on soil biota and processes as the fail safe number was 241 548, which should be larger than 700 (see methods for explanation). Moreover, funnel plots 242 showed skewed data (supplement Fig. S1), which suggest that positive effect sizes are more 243 likely to have been published than negative or neutral effect sizes. 244

245

246 *Feedbacks of exotics and natives from their own soil*

Overall, plant species experienced neutral plant-soil feedbacks from their own soil (summary 247 effect size: -0.008; 95 % bootstrap CI -0.16 to 0.14). However, the direction of effect sizes was 248 variable between studies ($Q_{total} = 274$, P = 0.001, d.f. = 208), meaning that plant species 249 250 experienced negative, positive and neutral effects from their own soil. Part of this variation was explained by the difference in feedback response between exotic and native species (see plant 251 origin effect in Table S5). Exotics had positive feedback when grown in soil conditioned by 252 themselves, whereas natives had negative feedback in their own soil (Fig. 3). However, plant 253 origin (native versus exotic species) did not explain all the variation in the effect sizes (Q_{within} 254 in Table S5). Interestingly, life forms tended to explain a portion of the variation in the effect 255 sizes ($Q_{between} = 7.62$, P = 0.054, d.f. = 3), but not all variation in effect sizes ($Q_{within} = 261$, P =256 0.001, d.f = 261). 257

Origin effects (native versus exotic) differed by plant life form (Table S5). Native grasses had negative feedback effects in their own soil, whereas exotic grasses had neutral feedback effects in soil conditioned by themselves (Fig. 3). Native forbs had negative feedback effects from soil conditioned by themselves, whereas exotics had positive feedback effects in

Meisner et al.

their own soil (Fig. 3). In contrast, native trees had positive feedbacks, whereas feedback effects of exotic trees were neutral (Fig. 3). The types of biomass measurement (aboveground biomass, total biomass, or other biomass measure) did not explain differences in effect sizes $(Q_{between} = 1.06, P = 0.59, d.f. = 1)$. The type of away soils (sterilized, or conditioned by other species or congener) did explain the differences between effect sizes (see Fig. S2; $Q_{between} =$ 11.57, P = 0.02, d.f = 4). There is no evidence for publication bias as the overall mean effect size was close to zero.

269

270 Feedbacks of exotics to natives and natives to exotics

271 Overall, native and exotic plant species experienced neutral feedback effects in soil conditioned by plants of the other origin (0.08; 95% CI -0.29 to 0.42). However, the direction 272 of the effect sizes varied between species ($Q_{total} = 200, P < 0.001, d.f. = 88$), with positive, 273 negative or neutral effect sizes all occurring. Origin (exotic versus native) did not explain the 274 heterogeneity among effect sizes (Fig. 4). Interestingly, life form explained part of the variation 275 276 among the effect sizes ($Q_{between} = 16.5$, P = 0.003; d.f. = 4), but not all ($Q_{within} = 181$, P < 0.001, d.f. 82). Moreover, exotic and native species responded differently to each others' soil within 277 plant life form (Table S6, Fig. 4). Changes in soil induced by exotic plant species did not 278 279 inhibit native species, except for native trees (Fig. 4). Interestingly, native grasses received a positive feedback from soil conditioned by exotics (Fig. 4, Table S6). Native and exotic forbs 280 experienced a neutral plant-soil feedback from each others' soil. Native trees and other life 281 forms experienced negative plant-soil feedback from exotics, whereas exotics had positive 282 feedbacks from natives (Fig. 4, Table S6). The biomass measurement method or type of soil 283 input (rhizosphere or litter) did not explain the variation between effect sizes (Qbetween for 284 biomass measurement method = 4.07, P 0.131, d.f. = 2; $Q_{between}$ for soil input = 1.22, P = 0.545, 285 d.f. = 2). There is no evidence for publication bias as the overall mean effect size was close to 286 287 zero.

289 **Discussion**

290 *Effects of exotics on soil biota and processes*

291 Our analysis revealed that effects of exotic species on soil biota and processes were neutral for most measurements, whereas exotics enhanced numbers of nematodes, invertebrates and C 292 cycling. These results confirm the results of a recent meta-analysis based on studies that had 293 taken an observational approach (Vilà et al. 2011). A potential problem of observational data is 294 that causes and consequences may not be separated. The advantage of our use of experimental 295 studies is that treatment effects point at causality, but a potential problem is that short duration 296 297 experiments may not fully reveal long-term processes, such as influences of exotic plant species on decomposers and decomposition. Therefore, we suggest that both approaches may 298 complement each other in providing a more complete insight in impacts of invasive exotic 299 plant species on community dynamics and ecosystem properties. 300

Our finding that exotic species increased C cycling is consistent with what has been 301 302 observed for invasive exotic and abundant native species (Liao et al. 2008), suggesting that those effects might be related to abundance rather than origin. However, increased C cycling 303 may be a consequence of trait differences between invasive exotic and native species, because 304 305 invasive exotic species often have traits associated with greater performance such as higher growth rates (Van Kleunen et al. 2010). In addition, invasive exotic plant species often have 306 higher nutrient concentrations in shoots and higher litter quality than native species (Agrawal 307 et al. 2005; Kurokawa et al. 2010; but see Godoy et al. 2010). These traits could contribute to 308 faster C cycling via for example enhanced decomposition rates (Cornwell et al. 2008). 309

In spite of effects of exotic plant species on some soil properties, it still remains an open question as to what extent changes in soil biota may be responsible for invasiveness of the exotic species. For example, the direction of effects and the effect sizes of the exotic plant species on soil characteristics did not match well with the observed plant-soil feedback

Meisner et al.

responses. This suggests that the drivers of invasiveness of exotic plant species are not 314 necessarily found among the changed soil parameters, but rather in subtle shifts in soil 315 microbes and processes (Inderjit and van der Putten 2010). Alternatively, it may be possible 316 317 that conditions present in the environment prior to the establishment of an exotic species may be important in contributing to the subsequent success of exotic plant species. As such the 318 'vacant niche hypothesis' suggests that certain exotic species may become successful because 319 320 they have access to resources in their new community that the native species do not use (Hierro et al. 2005). For example, disturbances such as N deposition in N poor ecosystems may 321 promote invasion of exotic species if the natives are not adapted to high N availabilities (Weiss 322 323 1999; Huenneke et al. 1990).

324

325 *Feedbacks of exotics and natives from their own soil*

The results supported part of our hypothesis in that overall exotics have positive plant soil-326 feedbacks in their own soil, whereas natives have negative feedback effects in their own soil. 327 This result is in contrast with a previous meta-analysis where both exotics and natives had 328 overall negative feedback effects in their own soil (Kulmatiski et al. 2008). This difference 329 may be due to the recent increase in studies with plant-soil feedback for exotic species 330 331 resulting in a larger sample size of 19 papers. However, opposite to our prediction, exotics did not have less negative plant-soil feedbacks than natives when analyzing the data across all life 332 forms. This turned out to be due to a difference between grasses and forbs on the one hand and 333 trees on the other. Native grasses and forbs experienced negative plant-soil feedbacks, whereas 334 soil feedback effects of native trees were positive. That finding is in line with a previous meta-335 analysis where trees had more positive feedbacks with their soil than forbs and grasses 336 (Kulmatiski et al. 2008). Our data suggest that this does not apply to exotic grasses, forbs and 337 trees as grasses and trees had neutral feedbacks with their soil while forbs had positive 338 feedbacks. Therefore, when comparing feedback effects of exotics with natives, care should be 339

Meisner et al.

taken to ensure proper comparisons, such as within life forms. Moreover, future experiments 340 may enhance understanding of invasiveness when they include factors, such as successional 341 position (Kardol et al. 2006), or time since introduction (Diez et al. 2010), which allows to 342 343 study why invasiveness of exotic species declines over time (Simberloff and Gibbons 2004). In their own soil, all life forms of exotic species had neutral feedback effects, whereas 344 native grasses had negative feedback (Fig. 3). This could be due to a lack of specialist 345 pathogens and less dependence on specialist mutualists for exotic plant species (Van der Putten 346 et al. 2007). For example, part of the success of *Prunus serotina* as an invader in Europe is 347 because virulent soil pathogens that keep this plant in check in its native range in the USA 348 appear to be absent in the invaded range (Reinhart et al. 2010). In most studies, data on 349 pathogen species and their virulence, however, are not available. Another explanation for the 350 neutral plant-soil feedback effect of exotics may be that their dependence on belowground 351 symbiotic mutualists is lower than for native plant species (Seifert et al. 2009; Vogelsang and 352 Bever 2009). Also in the case of symbiotic mutualists, data on community composition and 353 354 effects on plant performance are too rare for inclusion in a meta-analysis.

355

356 *Effects from exotics to natives and natives to exotics*

357 The soil feedback effects of exotics on natives depended on plant life form. Native trees experienced overall negative feedbacks from soil conditioned by exotic species, whereas native 358 grasses experienced positive feedbacks from exotics (Fig. 4). These conclusions appear to be in 359 contrast with a recent meta-analysis showing that soil from exotic species had a negative effect 360 on native species in comparison to performance in their own soil (Suding et al. 2013). Different 361 362 inclusion criteria may have been a reason for the discrepancies between these two studies. While we included studies with feedback effects of exotics to natives or natives to exotics, 363 Suding et al. (2013) had a smaller subset to work with because of including only those studies 364 that reported both feedback effects from exotics to natives and vice versa. Moreover, in our 365

Meisner et al.

study some older papers have been included further enlarging our dataset. Our result confirms
the suggestion (Suding et al. 2013) that plant life form may matter for plant-soil feedbacks of
exotic species to natives.

369 Soil conditioned by exotics had a positive effect on native grasses, which may result from effects of the exotic species on microbial community composition (e.g. Hawkes et al. 370 2006; Kourtev et al. 2003; Morriën and van der Putten 2013) and by increased faunal 371 abundance and C cycling (Fig. 2). The literature contains some examples of individual exotic 372 species that inhibit native species via the accumulation of local pathogens (Eppinga et al. 2006; 373 Mangla et al. 2008) or through inputs of novel allelochemicals into the soil that inhibit native 374 375 trees by a reduction in their symbiont (Stinson et al. 2006). However, our results do not suggest that exotic plant species in general inhibit native plant species via altered plant-soil feedbacks. 376 Overall exotic species had neutral feedback effects in soil conditioned by natives. This 377

conditioned by native trees. These results would suggest that exotic trees might benefit from (at
least some) of the symbiotic mutualists of the native tree species as suggested previously
(Richardson et al. 2000; Gundale et al. 2014).

applied to most plant life forms, except that exotic trees had positive feedback effects in soil

382

378

383 Conclusion

Our results show that plant life form matter when studying potential effects of exotic invaders on native plant community composition. Exotic species may promote native grasses, whereas they may inhibit native trees. Therefore, we suggest that when assessing effects of exotic plant species on subsequent potential establishment of native plant species (Grman and Suding 2010), effects of plant life form need to be explicitly taken into account. Our results suggest that plant life forms are not only important to consider when comparing plant traits (Tecco et al. 2010), but also when comparing plant-soil feedback effects between native and exotic plant

Meisner et al.

species. Further studies might as well reveal other traits that relate to patterns in plant-soilfeedback effects of native and exotic plant species.

The observed feedback responses of exotics and natives could not be related directly to their influences on general soil biotic and abiotic characteristics. Therefore, more subtle effects on soil conditions, such as the population abundance of specialist pathogens and symbionts, may need to be addressed. In addition, the effects of changed composition of these soil specialists on the performance of native and exotic plant species need to be quantified in order to further understand the observed patterns in plant-soil feedback effects.

399

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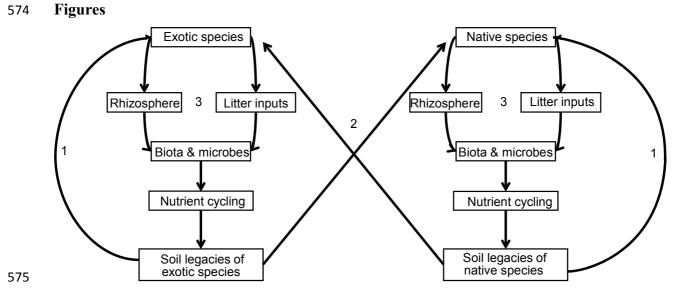


Fig. 1 Schematic overview of plant-soil feedback. Exotic and native plant species can influence
soil legacies via rhizosphere and/ or litter inputs. These soil legacies can feedback to influence
the performance of the plant species that caused the change in the soil properties (1) and the
performance of a neighbouring plant (2). In the present study, we have compared plant-soil
feedbacks within and between native and exotic plant species. We further have studied effects
of exotic species on soil biota, microbes and soil processes (3). Scheme is adapted from Bever
(Bever 2003; Bever et al. 1997).

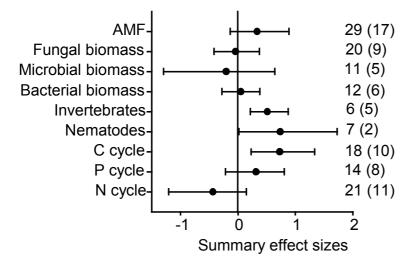
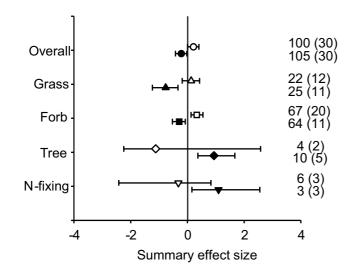


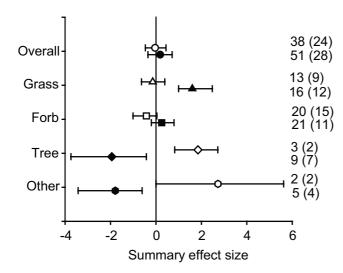
Fig. 2 Effects of exotics on soil biota and processes. Effect sizes of effects from exotics on soil variables were calculated as the difference between soil conditioned by the exotic species and soil conditioned by the native species. Confidence intervals that do not overlap 0 indicate that exotics had an effect on the soil parameter at P < 0.05. Number of species and in brackets numbers of studies are presented on the right side of the graph.





592

Fig. 3 Plant-soil feedbacks of home soil for exotics (open circles) versus natives (closed 593 circles). Effect sizes were calculated by the difference between soil conditioned by own plant 594 species (home soil) and sterilized soil or soil conditioned by other plant species (away soil). 595 Exotics differed from natives for overall feedback (circles), grasses (up facing triangle), forbs 596 (squares) and trees (diamond), but not for N-fixing species (down-facing triangle) (Qbetween 597 Table 1). Bootstrap bias corrected 95 % confidence intervals that do not overlap 0 indicate 598 positive or negative plant-soil feedbacks at P < 0.05. Number of species and numbers of studies 599 (in brackets) are presented on the right side of the graph. 600 601



602

Fig. 4 Plant-soil feedbacks of natives to exotics (open circles) versus exotics to natives (closed 603 circles). For natives, effect sizes were calculated by the difference between soil conditioned by 604 the exotic (away soil) and soil conditioned by themselves (home soil). For exotic, effect sizes 605 606 were calculated by the difference between soil conditioned by the native species (away soil) and soil conditioned by themselves (home soil). Exotics and natives differed in response to 607 each other's plant-soil feedback for grasses (up-facing triangle), trees (diamond) and other 608 species (hexagon), but not for overall (circles) and forbs (square) (Table 2, Q_{between}). Bootstrap 609 corrected 95 % confidence intervals that do not overlap 0 indicate positive or negative plant-610 soil feedbacks at P < 0.05. Number of species and number of studies (in brackets) are 611 presented on the right side of the graph. 612